



# Ecology Assessment

## Ayrburn Screen Hub

### Waterfall Park Developments Limited

c/o Winton Group Holdings Limited, PO 105526, Auckland 1143

Prepared by:

**SLR Consulting New Zealand**

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Revision	Date	Prepared By	Checked By	Authorised By
1.0	25 July 2025	Ben Ludgate	Fleur Tiernan	Ben Ludgate

## Basis of Report

This report has been prepared by SLR Consulting New Zealand (SLR) with all reasonable skill, care and diligence, and taking account of the timescale and resources allocated to it by agreement with Waterfall Park Developments Limited (the Client). Information reported herein is based on the interpretation of data collected, which has been accepted in good faith as being accurate and valid.

This report is for the exclusive use of the Client and may be disclosed by the Client to the Environmental Protection Authority (EPA) and the EPA Hearing Panel as part of the Client application. No warranties or guarantees are expressed or should be inferred by any third parties. This report may not be relied upon by other parties without written consent from SLR.

SLR disclaims any responsibility to the Client and others in respect of any matters outside the agreed scope of the work.



## Executive Summary

Waterfall Park Developments Limited (WPDL) is proposing to develop a screen-production facility with associated accommodation, collectively referred to as the Screen Hub, at Ayrburn Farm, Arrowtown. WPDL engaged SLR to provide an assessment of ecological effects (AEcE) as part of WPDL's permission application for the development through a substantive application under the Fast-track Approvals Act 2024. The report produced by SLR may be disclosed by the Client to the EPA and the EPA Hearing Panel as part of the Client application.

The Screen Hub includes workshops/work rooms, offices, dressing rooms, backlot space, and accommodation units with landscaping, parking, and associated infrastructure. Stormwater facilities include a network of sediment retention ponds, rain gardens, engineered wetlands, and a treatment pond that will regulate sediment and water run-off from the Screen Hub area. The proposed construction area will be located on land adjacent to Mill Creek. WPDL also propose the construction of an in-line sediment trap in Mill Creek to help with improving the water quality of Lake Hayes and the downstream sections of Mill Creek.

This report provides an AEcE for the development. The assessment is mainly based on existing data available for Mill Creek. Water quality has been sampled monthly since November 2018 at three sites: Upstream, adjacent to (i.e., "Boundary"), and Downstream of the proposed Screen Hub area by WPDL. In addition, the Otago Regional Council (ORC) monitor water quality in Mill Creek approximately 1.5 km downstream of the proposed Screen Hub area (the Fish Trap site).

Annually in May, macroinvertebrate surveys have been conducted at each of the four water quality sites (Upstream, Boundary, Downstream, Fish Trap), with substrate surveys also undertaken at eight locations between the Downstream and Upstream sites.

Monitoring shows that the current state of the aquatic ecosystem in Mill Creek is of good quality. Water quality data shows some water quality metrics were highest around 2022, but most metrics have decreased back to levels recorded in 2019. The exceptions are nitrogen compound concentrations, which show an increasing trend in concentrations. The increasing trend is present for both the Upstream and Downstream sites and is therefore related to wider upper catchment activities and not localised within the development area or related to the construction activities in Waterfall Park. Notably, for the 2024 and 2025 sampling, all water quality metrics met relevant Regional Plan: Water for Otago limits except nitrate nitrite nitrogen limits. Mill Creek supports a healthy macroinvertebrate community and adequate feeding, refuge, and breeding grounds for a relatively large trout population.

Potential effects from the development that may adversely impact water quality and ecological values of Mill Creek are:

- Sediment discharge during construction.
- Contaminant spills and pest introduction.
- Concrete.
- Stormwater discharge post development.
- Sediment trap construction and maintenance.

Mitigation to avoid and minimise any adverse effects include sediment and erosion control methods to manage sediments during construction works, such as silt fences, sediment retention ponds, clean water diversion channels, and dirty water diversion channels. The



bulk earthworks associated with the sediment trap are proposed for spring and summer when periods of fine weather are more frequent.

Mitigation to avoid contaminant spills and pest introduction include storage of chemicals and fuels will be as far as practicably possible from waterways, onsite refuelling of vehicles and plant machinery will occur in a designated refuelling bay, and wastes will be stored safely until either recycled, reused, or disposed of safely offsite. Weeds will be treated prior to disturbance of the natural surface and weed free topsoil will be retained for reuse in site rehabilitation.

Mitigation to avoid adverse effects from concrete works include work being undertaken in dry weather and confined to dry riverbeds to prevent materials from entering water systems and all concrete washing being undertaken in a designated concrete wash-out pit.

Stormwater management for the Screen Hub area includes a network of sediment retention ponds, rain gardens, engineered wetlands, and a treatment pond that will regulate sediment and water run-off from the area.

Whilst the creation of the in-line sediment trap will result in disturbance of the stream bed, the construction will ensure that fish are not adversely affected through their removal and relocation prior to works commencing. Fish passage will not be impeded for the duration of the construction, which will be completed within three months, outside of the fish spawning season and at low flow levels.

The development will have an overall positive effect on the environment. Water quality in the lower Mill Creek reaches and in Lake Hayes will improve through the removal of contaminants such as fine sediment and phosphorus in the sediment trap and the change in land use which is predicted to reduce nitrogen and phosphorus loads to the creek and Lake Hayes. By addressing key sources of phosphorus and sedimentation, these initiatives align with initiatives already in place to support the long-term goal of improving the water clarity and ecological balance of Lake Hayes.



## Table of Contents

<b>Basis of Report .....</b>	<b>i</b>
<b>Executive Summary .....</b>	<b>ii</b>
<b>Acronyms and Abbreviations .....</b>	<b>vi</b>
<b>1.0 Introduction .....</b>	<b>7</b>
<b>2.0 Assessment Methodology .....</b>	<b>7</b>
<b>3.0 Existing Environment.....</b>	<b>9</b>
3.1 Mill Creek Catchment .....	9
3.2 Water Quality and Flow Characteristics .....	10
3.3 Instream Substrate .....	16
3.4 Periphyton .....	18
3.5 Benthic Macroinvertebrates .....	18
3.6 Fish .....	21
3.7 Summary.....	22
<b>4.0 Assessment of Actual and Potential Effects .....</b>	<b>23</b>
4.1 Project Components .....	23
4.2 Sediment Discharge During Construction .....	23
4.3 Contaminant Spills and Pest Introduction .....	24
4.4 Concrete .....	24
4.5 Stormwater Discharge Post Development .....	25
4.6 Sediment Trap Construction and Maintenance .....	25
<b>5.0 Monitoring.....</b>	<b>30</b>
<b>6.0 Positive Effects.....</b>	<b>30</b>
<b>7.0 References.....</b>	<b>33</b>
<b>8.0 Closure.....</b>	<b>34</b>

## Figures in Text

- Figure 1: Map of Mill Creek and the ephemeral tributary in the Ayrburn area, with proposed Screen Hub boundary outlined in orange. Water quality monitoring sites are shown as blue points. .... 8
- Figure 2: Mill Creek at Fish Trap mean daily flow (m<sup>3</sup>/s), 2018 to 2025. Water quality monitoring occasions indicated by black points. The dashed line shows the median flow (0.38 m<sup>3</sup>/s). .... 10
- Figure 3: Turbidity (NTU), recorded by continuous monitoring loggers installed at the Downstream and Upstream monitoring sites, in Mill Creek between 2019 and 2025. Note y-axes are truncated for some years (i.e., 2020, 2021). .... 12



Figure 4:	Long-term predicted trend ( $\pm 95\%$ confidence intervals) of physical and chemical water quality metrics at Mill Creek monitoring sites, 2019 to 2025. Dashed horizontal lines denote relevant RPW limits for <i>E. coli</i> (260 cfu/100 mL; 'a') and turbidity (5 NTU; 'd').	14
Figure 5:	Long-term predicted trend ( $\pm 95\%$ confidence intervals) of nutrient concentrations at Mill Creek monitoring sites, 2019 to 2025. Dashed horizontal lines denote relevant RPW limits for dissolved reactive phosphorus (0.01 mg/L; 'a'), nitrate+nitrite nitrogen (0.075 mg/L; 'b'), total ammoniacal nitrogen (0.1 mg/L; 'c').	15
Figure 6:	Map of the sediment sampling locations (numbered red points) and reaches used in bankside sports fish and redd surveys (lettered sections in light blue) in Mill Creek.	17
Figure 7:	Composition of particle size classes in substrate samples across Mill Creek sites, May 2024. Site numbers in order from Downstream (1) to Upstream (8). Gravels suitable for trout spawning (size range 8 to 64 mm) are dotted.	18
Figure 8:	Long-term macroinvertebrate metrics across Mill Creek sites. Regression lines represent a line of best fit but may not accurately reflect trends due to the sample size being too small for reliable statistical analysis and when there is high metrics variability.	20
Figure 9:	Number of brown trout (a) and redds (b) observed across Mill Creek reaches, 2024. Reaches in order from downstream (Reach A) to upstream (Reach H) (see Figure 6).	22
Figure 10:	Schematic of the proposed sediment trap. Prepared by Patersons.	26
Figure 11:	Examples of existing sediment traps located in the upper Mill Creek catchment. Top photo from Friends of Lake Hayes. Bottom photo from Mana Tāhuna Charitable Trust (2023).	29

## Appendices

<b>Appendix A</b>	<b>Authors' Qualifications and Experience</b>
<b>Appendix B</b>	<b>Mill Creek Water Quality Site Photographs</b>
<b>Appendix C</b>	<b>Mill Creek Macroinvertebrate Community, 2024</b>
<b>Appendix D</b>	<b>Erosion and Sediment Control Plan Drawing (from EMP)</b>



## Acronyms and Abbreviations

AEcE	Assessment of Ecological Effects
ARI	Average Recurrence Interval
CFFA	Complex Freshwater Fisheries Activities
DO	Dissolved oxygen
<i>E. coli</i>	<i>Escherichia coli</i>
eDNA	Environmental DNA
EMP	Environmental Management Plan
EPA	Environmental Protection Authority
EPT	Ephemeroptera (mayflies), Plecoptera (stoneflies), and Trichoptera (caddisflies)
FRE3	Flow events that exceed three times the median flow
LAWA	Land, Air, Water Aotearoa
MCI	Macroinvertebrate Community Index
NTU	Nephelometric Turbidity Unit
NZFFD	New Zealand Freshwater Fish Database
NZTCS	New Zealand Threat Classification System
ORC	Otago Regional Council
REC	River Environment Classification
REL	Ryder Environmental Limited
RPW	Regional Plan: Water for Otago
SFFA	Standard Freshwater Fisheries Activities
SMP	Stormwater Management Plan
SOE	State of the Environment
SQMCI	Semi Quantitative Macroinvertebrate Community Index
WPD	Waterfall Park Developments Limited



## 1.0 Introduction

Through a substantive application under the Fast-track Approvals Act 2024, Waterfall Park Developments Limited (WPDL) is proposing to develop a screen-production facility with associated accommodation, collectively referred to as the Screen Hub, at Ayrburn Farm, Arrowtown. The Screen Hub includes workshops/work rooms, offices, dressing rooms, backlot space, and accommodation units with landscaping, parking, and associated infrastructure (i.e., road, stormwater facilities). The stormwater facilities include a network of sediment retention ponds, rain gardens, engineered wetlands, and a treatment pond that will regulate sediment and water run-off from the Screen Hub area. The proposed construction area will be located on land adjacent to Mill Creek and an unnamed ephemeral tributary of Mill Creek (Figure 1). This proposal also includes an initiative to add an in-line sediment trap (50 m long by 12 m wide) to the lower Mill Creek section within the proposed construction area, which is designed to assist with improving the water quality of Lake Hayes and downstream sections of Mill Creek. The sediment trap will capture significant amounts of sediment being transported in the creek and allow the extraction of this sediment from the system, thereby removing the sediment and trapped nutrients before they could enter Lake Hayes and contribute to water quality issues, such as cyanobacteria blooms.

WPDL engaged SLR to provide an assessment of ecological effects (AEcE) as part of WPDL's permission application for the development through a substantive application under the Fast-track Approvals Act 2024. The report produced by SLR may be disclosed by the Client to the EPA and the EPA Hearing Panel as part of the Client application. This AEcE aims to describe the current ecological characteristics of the area including water quality, macroinvertebrates, aquatic plants, and fish communities and discusses the potential effects of the proposal on the existing environment, with reference to the associated Environmental Management Plan (EMP; Enviroscope 2025) and Stormwater Management Plan (SMP; CKL 2025).

## 2.0 Assessment Methodology

This AEcE is a desktop review primarily based on ongoing data collections conducted in and around Mill Creek, outlined below:

Water quality has been sampled monthly since November 2018 at three sites: Upstream, adjacent to (i.e., "Boundary"), and Downstream of the proposed Screen Hub area by WPDL (Figure 1; see photos in Appendix B). In addition, continuous loggers for turbidity have been established at the Upstream and Downstream sites. Otago Regional Council (ORC) has a long-term State of the Environment (SOE) monitoring site in Mill Creek (Fish Trap) approximately 1.5 km downstream of the proposed Screen Hub area, close to where the creek enters Lake Hayes (Figure 1).

Annually in May, macroinvertebrate surveys have been conducted at each of the four water quality sites (Upstream, Boundary, Downstream, Fish Trap), with substrate surveys also undertaken at eight locations between the Downstream and Upstream sites. Observational surveys of sports fish and 'redds' (spawning patches) have also been conducted on several occasions annually during autumn between the Downstream and Upstream sites.







**Figure 1: Map of Mill Creek and the ephemeral tributary in the Ayrburn area, with proposed Screen Hub boundary outlined in orange. Water quality monitoring sites are shown as blue points.**



## 3.0 Existing Environment

### 3.1 Mill Creek Catchment

#### Mill Creek

Mill Creek is largely spring-fed and has a total catchment area of approximately 55 km<sup>2</sup>. The stream is categorised as a cool dry, hill, hard sedimentary, pastoral, mid order high gradient stream (CD/H/HS/P/MO/HG) under the New Zealand River Environment Classification (REC) system. The Mill Creek catchment drains into Lake Hayes, which is approximately 1.5 km downstream of WPDL's proposed development. Historically, much of the land use within this catchment consisted primarily of cattle and sheep grazing on exotic pasture, as well as a significant amount of urban development including golf courses. However, in recent years, agricultural land use has been replaced with ongoing and planned urban development. Lake Hayes has likely undergone progressive eutrophication (nutrient enrichment) since development and intensification began in the catchment. Consequently, efforts have been made to better understand and improve the water quality of the lake, including the preparation of the 'Lake Hayes Restoration and Monitoring Plan' (Hydrosphere Research 2017).

Since 2019, Mill Creek has undergone considerable change in association with the staged construction of the Waterfall Park development. The works involved the installation of six weir structures, seven single-span bridges and two culvert crossings on, in, and over the bed of Mill Creek. Additionally, the project included the widening and reshaping of Mill Creek, and the temporary diversion and re-instatement of the creek to/from a temporary channel so works could be undertaken in the dry creek channel. Extensive riparian enhancement (fencing and planting) of the creek has also been completed as part of the Waterfall Park development.

#### Ephemeral Mill Creek tributary

An unnamed ephemeral spring-fed tributary of Mill Creek is located on the western side of Ayrburn Farm (Figure 1). The ephemeral watercourse is small (order 1), with a catchment area of approximately 0.43 km<sup>2</sup>. Observations made on site suggest a typical flow to 0-2 L/s, with a predicted 100-year average recurrence interval (ARI) of 750-900 L/s (Fluent Solutions 2020).

Observations of the tributary from 2020 (reported by Ryder Environmental Limited (REL) 2020) include the wetted length of the ephemeral watercourse ranging between 300-800 m long. The tributary initially runs from north to south before turning towards the south-east and entering Mill Creek at the downstream boundary of Ayrburn Farm. The channel is mostly poorly defined, being located within a depression and without clear banks, and is only connected to Mill Creek during flood conditions. In places, there are small areas of open shallow water, some that have been formed artificially by the construction of weirs. It is likely that these areas would also dry up at times. For much of its length the channel is open to stock grazing and is dominated by pasture grasses. The south-eastern section of the channel is overgrown with willow, pine trees and other exotic vegetation, and in the lower area of the tributary near Mill Creek there is no identifiable channel.

Due to its lack of permanent flow the ephemeral watercourse does not provide any habitat for fish and only provides minimal habitat for other aquatic life. Benthic macroinvertebrates were sampled in May 2020 and the community was dominated by snails, worms, crustaceans, and fly larvae, indicative of 'poor' water quality and/or habitat condition (according to Macroinvertebrate Community Index (MCI) and Semi Quantitative MCI





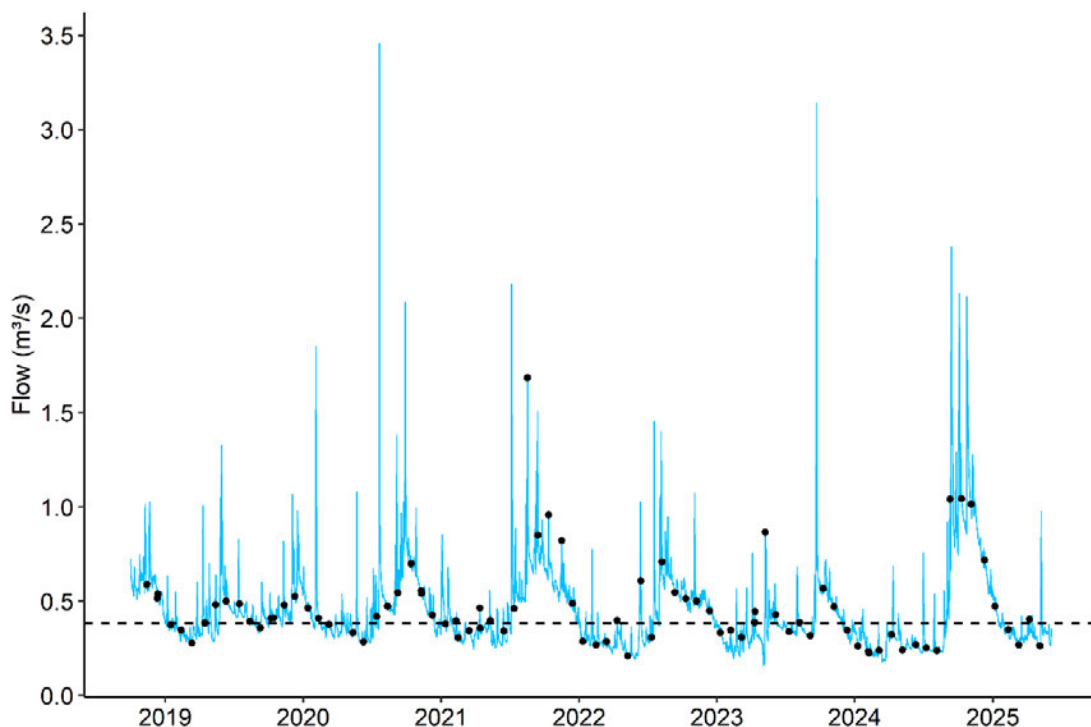
(SQMCI) scoring; Stark 1998). The modified, non-permanent nature of the habitat means that any aquatic life present is not of high significance (REL 2020).

### 3.2 Water Quality and Flow Characteristics

Water quality monitoring has been generally completed on a monthly basis at all four water quality sites since October 2018. Water quality descriptions discussed below reflect trends observed over the six and a half year period (2018-2025). Water quality limits for Mill Creek, from Schedule 15 of ORC's Regional Plan Water for Otago (RPW), are discussed below and are denoted in figures where appropriate. The water quality limits are set for the receiving environment and are based on achieving Good Quality Water for Mill Creek. Specifically, the RPW notes that these limits are achieved when 80% of samples collected at a site, when flows are at or below median flow, over a rolling 5-year period, meet or are better than the limits in Schedule 15 (Table 15.2.2 in RPW, 2022). All available water quality data has been presented here, including data when flows were above the median flow, and therefore comparisons with RPW limits are indicative only.

#### Physical and Chemical

Over the past seven and a half years, Mill Creek mean daily flows (monitored at the Fish Trap site) ranged between 0.16 and 3.46 m<sup>3</sup>/s with a median of 0.38 m<sup>3</sup>/s (Figure 2); instantaneous peak flows over this period have ranged from 0.11 to 6.18 m<sup>3</sup>/s. According to New Zealand River Maps (Whitehead and Booker 2020), Mill Creek has an average of ten flow events per year that exceed three times the median flow (i.e., "FRE3"), which provides an indication of 'flashiness'. For comparison, the nearby Arrow River, which has a much larger catchment than Mill Creek, has only seven FRE3 events per year.



**Figure 2: Mill Creek at Fish Trap mean daily flow (m<sup>3</sup>/s), 2018 to 2025. Water quality monitoring occasions indicated by black points. The dashed line shows the median flow (0.38 m<sup>3</sup>/s).**

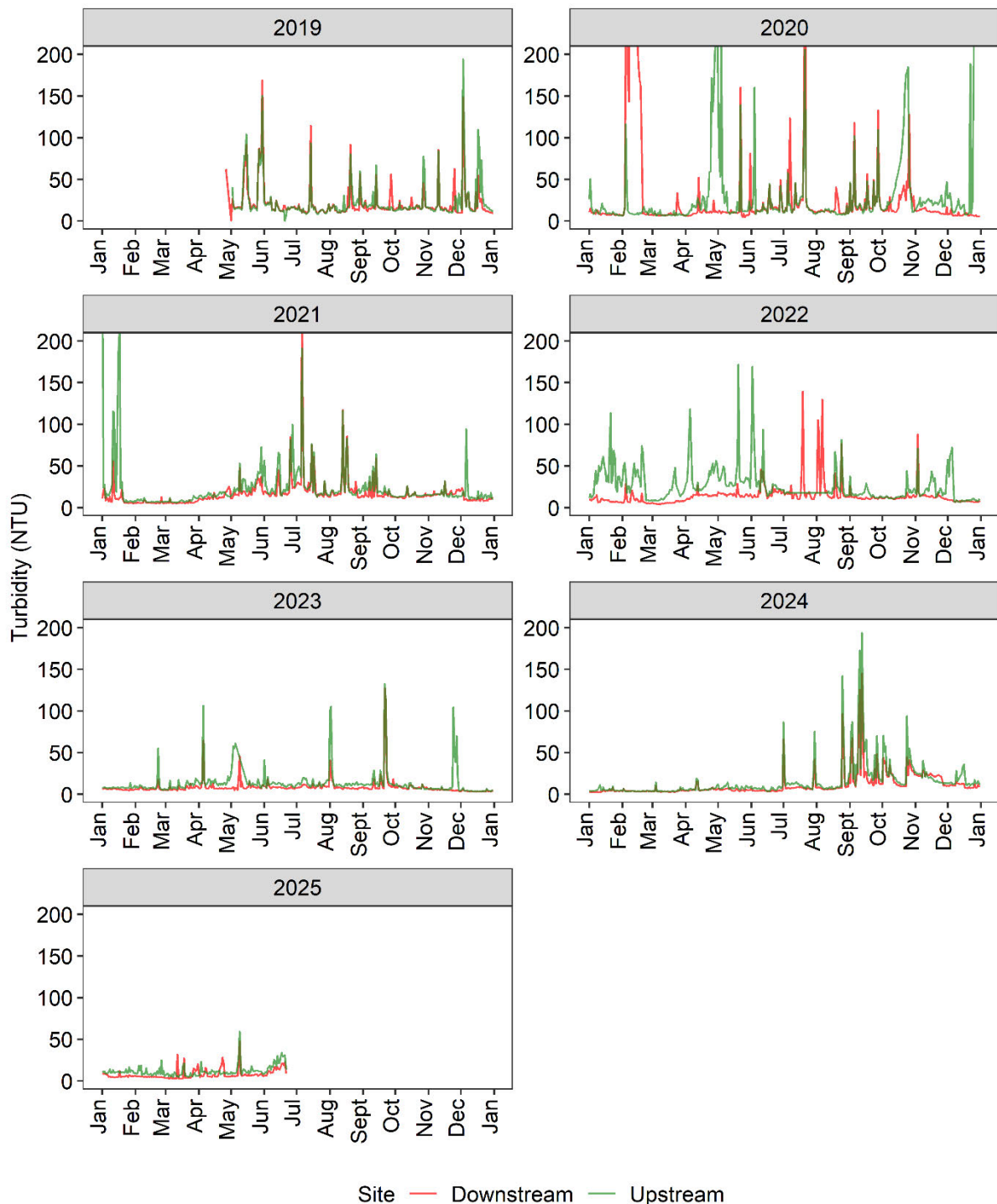


Continuous monitoring of turbidity, measured in Nephelometric Turbidity Units (NTU), has found that turbidity readings were generally similar between Upstream and Downstream sites. Since 2019 the Downstream site often had marginally lower turbidity and fewer and lower spikes (Figure 3). Overall, in addition to some site-specific variation, there has been substantial temporal variation between years and months, with 2023, 2024 and 2025 having the lowest turbidity overall.

CKL (2024) analysis of the turbidity data has found that in 2019, prior to construction activities in the creek, the turbidity levels at the Downstream site were more often worse than Upstream, while in recent years there has been a higher percentage of times when turbidity readings were better at the Downstream site than at the Upstream site. However, a comparison of 'better' or 'worse' values must be interpreted carefully as this does not incorporate the intensity of specific differences between sites. Overall, while turbidity at the Downstream site is often similar to, or marginally lower than, that observed at the Upstream site (Figure 3), there is no consistent evidence to suggest that turbidity at the Downstream site is of 'better' quality overall. What is evident from the data is that although notable peaks in turbidity have still occurred at each site in recent years, the synchronicity in turbidity fluctuations at the Upstream and Downstream sites indicate that sediments were entering the creek from the upper catchment, and there were no notable sources of new sediments entering the stream from within the construction site. Turbidity data from 2024 shows that the turbidity peaks in winter/spring are associated with higher flows, similarly the lower turbidity levels recorded for 2025 reflect lower flows, where instantaneous flows have been mostly below  $0.5 \text{ m}^3/\text{s}$ , with a peak flow of  $1.7 \text{ m}^3/\text{s}$  on 9 May 2025.

Dissolved oxygen (DO) has been measured in Mill Creek at the Fish Trap since May 2023 by dataloggers recording DO every five minutes. The data record shows the waters are well oxygenated, with DO saturation rarely going below 90%.





**Figure 3: Turbidity (NTU), recorded by continuous monitoring loggers installed at the Downstream and Upstream monitoring sites, in Mill Creek between 2019 and 2025. Note y-axes are truncated for some years (i.e., 2020, 2021).**



Comparisons of monthly sample data has found that the long-term trend for *Escherichia coli* concentrations at the Downstream and Boundary sites peaked around 2022 but have since decreased to below the RPW limit (Figure 4a). Similar peaks in *E. coli* occurred at the Fish Trap and Upstream sites, but no long-term trends are apparent due to concentrations being more consistent, with fewer high concentrations (spikes), and lower overall.

Total suspended solids at the Boundary and Downstream sites show a decreasing trend from approximately 2022 to 2025 whilst the Upstream site shows a somewhat increasing trend for the same time period (Figure 4c). Trends for turbidity are less clear, however Upstream turbidity shows a slight increasing trend from 2022 onwards (Figure 4d). Turbidity levels from 2023 onwards were mostly below the RPW limit for each site. Comparatively, pH levels have remained generally consistent over time (Figure 4b).

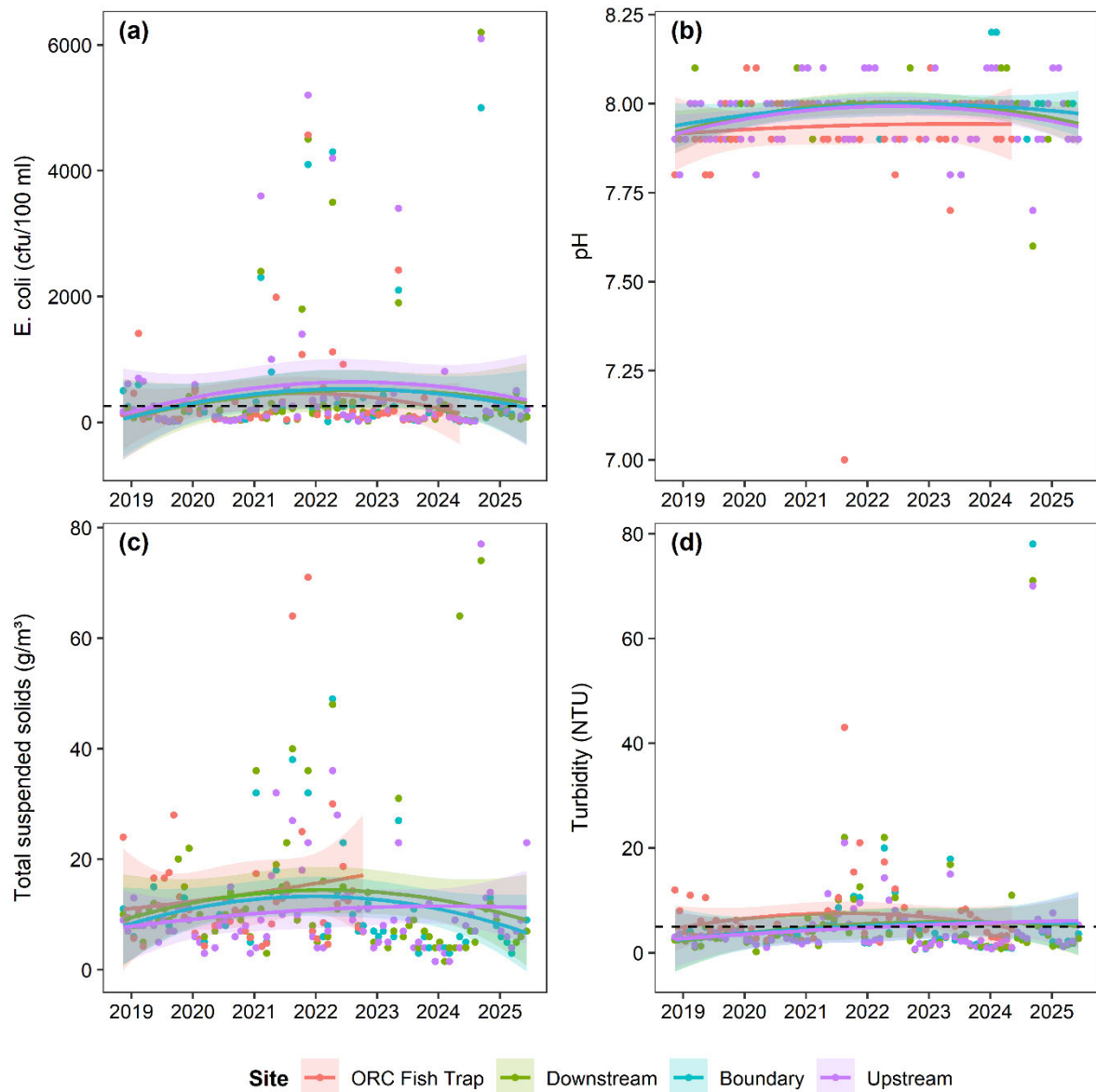
## Nutrients

Nitrate nitrite nitrogen and total nitrogen concentrations show increasing trends at all sites (Figure 5b and d). The concentrations decreased at first from 2019 to the lowest concentrations recorded in 2021, before increasing with the highest concentrations recorded in 2024/25. The trend is most apparent at the Upstream site, however concentrations were very similar at all three sites suggesting that the trend in concentrations is being driven by upstream activities. Nitrate nitrite concentrations at all sites, for the entire monitoring period, are above the RPW limit. Ammoniacal nitrogen concentrations show little trends over time but show a similarity with nitrate nitrite and total nitrogen concentrations with upstream ammoniacal nitrogen concentrations often higher than downstream concentrations (Figure 5c). Ammoniacal nitrogen concentrations have consistently been well below the RPW limit at all three sites.

Dissolved reactive phosphorous concentrations showed an increasing trend from 2019 to 2021/22, before decreasing again, with concentrations recorded in 2025 similar to those recorded in 2019 and below the RPW limit (Figure 5a). The temporal trend was apparent for Upstream, Downstream and Boundary sites. Total phosphorus concentrations showed a similar pattern (Figure 5e).

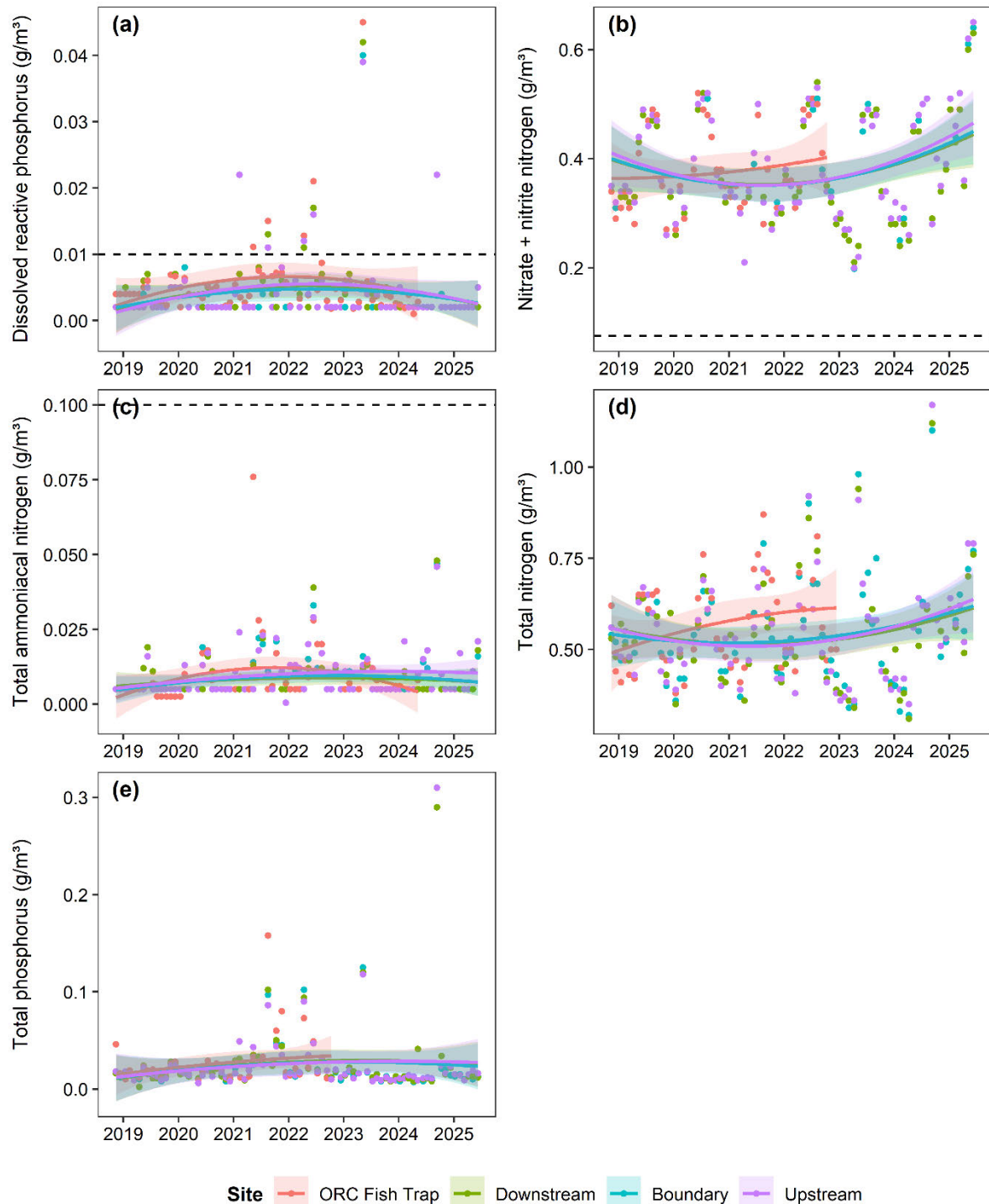
Long-term water quality monitoring at Mill Creek shows that while many metrics were highest around 2022, there has been a steady improvement since, supported by enhancements to stream infrastructure such as stabilised banks and riparian plantings. Water quality is expected to remain stable as a result of these interventions. However, nitrate nitrite nitrogen concentrations have remained consistently high across all sites since monitoring began, suggesting that these elevated nitrogen concentrations are likely attributed to groundwater and broader land use practices in the surrounding catchment, rather than construction activities related to the Waterfall Park development.





**Figure 4: Long-term predicted trend (±95% confidence intervals) of physical and chemical water quality metrics at Mill Creek monitoring sites, 2019 to 2025. Dashed horizontal lines denote relevant RPW limits for *E. coli* (260 cfu/100 mL; 'a') and turbidity (5 NTU; 'd').**





**Figure 5: Long-term predicted trend (±95% confidence intervals) of nutrient concentrations at Mill Creek monitoring sites, 2019 to 2025. Dashed horizontal lines denote relevant RPW limits for dissolved reactive phosphorus (0.01 mg/L; 'a'), nitrate+nitrite nitrogen (0.075 mg/L; 'b'), total ammoniacal nitrogen (0.1 mg/L; 'c').**





### 3.3 Instream Substrate

Substrate particle size distribution has been assessed on an annual basis in Mill Creek. As the 2024 sampling followed the completion of significant instream works (e.g., channel diversion, installation of instream structures), the 2024 results therefore indicate the expected longer-term substrate sizes. Substrates were highly diverse across the Mill Creek sites but were particularly so at sites near the Downstream monitoring site (substrate sites 1-3 in Figure 6) where most substrate classes were equally abundant (Figure 7). In contrast, sites further upstream, towards the Upstream monitoring site (substrate sites 6 and 8 in Figure 6), generally contained more large gravels and small cobbles. While there was some site-specific variation overall (e.g., median substrate size at Site 7 was comparatively lower than adjacent sites), these differences were likely underpinned by local stream structure and flow characteristics.

The preferred spawning habitat of brown trout is gravels within the size range 8 to 64 mm (i.e., small-medium to large gravels; Shirvell and Dungey 1983). Gravel substrates of this size were present in great proportions at most sites within Mill Creek in May 2024, but relatively less so at Sites 1, 2, 6, and 8; these substrate sites were located within observational reaches for trout where redd counts were found to be lowest overall.

Overall, the amount of fine clay/silt and sand particles at most sites was low (combined amount  $\leq 20\%$ ), which is favourable for successful egg incubation; too much fine sediment can cause smothering of eggs within redds. There has been no indication that instream works associated with instream or bankside developments have resulted in any increase in fine sediments within Mill Creek to date.

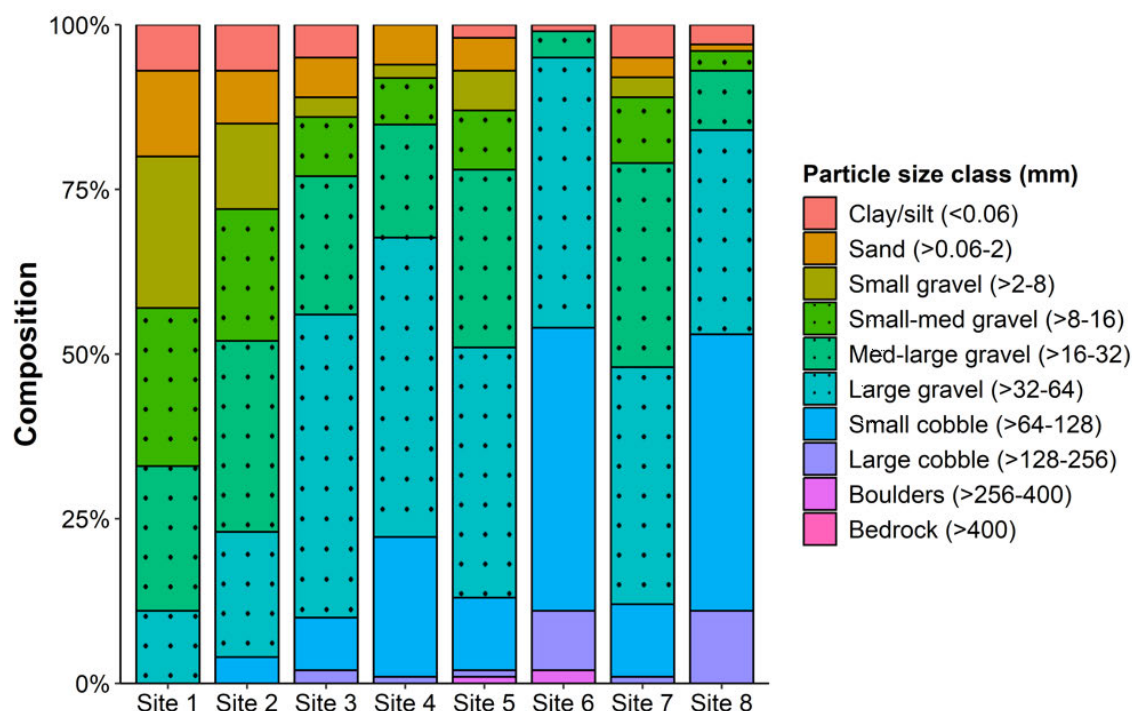




**Figure 6. Map of the sediment sampling locations (numbered red points) and reaches used in bankside sports fish and redd surveys (lettered sections in light blue) in Mill Creek.**







**Figure 7: Composition of particle size classes in substrate samples across Mill Creek sites, May 2024. Site numbers in order from Downstream (1) to Upstream (8). Gravels suitable for trout spawning (size range 8 to 64 mm) are dotted.**

### 3.4 Periphyton

Periphyton (algae) communities were sampled at each of the four Mill Creek water quality monitoring sites in January 2019 (REL 2020). Periphyton monitoring found that cover and biomass was typically low in Mill Creek within Waterfall Park, at the Boundary and Downstream sites. Higher periphyton levels were however observed further upstream, at the Upstream site. This difference was likely related to the variation in substrate between these sites, with the substrate upstream being dominated by larger and more stable substrates which are more favourable for periphyton growth than the smaller gravels found in downstream areas. Both mat and filamentous periphyton growth forms have been observed in Mill Creek, including the invasive algae *Didymosphenia geminata*.

Monitoring of periphyton communities was previously undertaken annually by the ORC at the Fish Trap site (discontinued in 2018). ORC periphyton biomass data from this site for 2011 to 2016 indicated 'poor' to 'fair' habitat quality (data sourced from the Land, Air, Water Aotearoa (LAWA) website; REL 2020), which is comparable to REL surveys.

No further sampling of periphyton in Mill Creek has occurred since 2019, however given the dominance of gravels throughout the creek, periphyton communities are expected to remain at low cover and biomass levels.

### 3.5 Benthic Macroinvertebrates

A total of 24 benthic macroinvertebrate taxa were identified across the four Mill Creek sites in the most recent 2024 sampling event, with 16 taxa being found at the Upstream site, 20



taxa at Boundary, 17 taxa at Downstream, and 12 taxa at Fish Trap (Appendix C). Despite this site-specific range in taxa richness, the number of sensitive Ephemeroptera (mayflies), Plecoptera (stoneflies) and Trichoptera (caddisflies) (EPT) taxa (excluding the more tolerant *Hydroptilidae*) was the same at all sites (six). However, the percentage of EPT taxa in invertebrate samples was higher at the Fish Trap site (50%) than other sites (30-38%). Overall, communities at each site were dominated by *Deleatidium* mayflies, *Pycnocentria* caddisflies, and *Potamopyrgus* snails. Following the scoring of Stark (1998), MCI scores at all sites were indicative of 'fair'/'poor-fair' habitat quality, however the more ecologically robust SQMCI scores indicated that habitat quality at the Fish Trap and Downstream sites was 'excellent', Boundary was 'good' and Upstream was 'fair-good'.

Long-term macroinvertebrate monitoring suggests that overall taxa richness and EPT taxa richness has decreased at the Fish Trap and Upstream sites, whereas the Boundary and Downstream sites have remained relatively stable or inconsistently variable (Figure 8a,b). Percent EPT taxa richness has also remained stable or variable at all sites except at Fish Trap, where percent EPT richness has increased in recent years (Figure 8c).

Long-term MCI scores have consistently remained indicative of 'fair' habitat quality at all sites (Figure 8d). In contrast, SQMCI scores have remained variable but 'fair' on average at the Upstream site and variable between 'good' and 'excellent' at the Boundary site. SQMCI scores at the Fish Trap and Downstream sites have substantially improved over recent years, going from 'fair-good' to 'excellent' (Figure 8e).

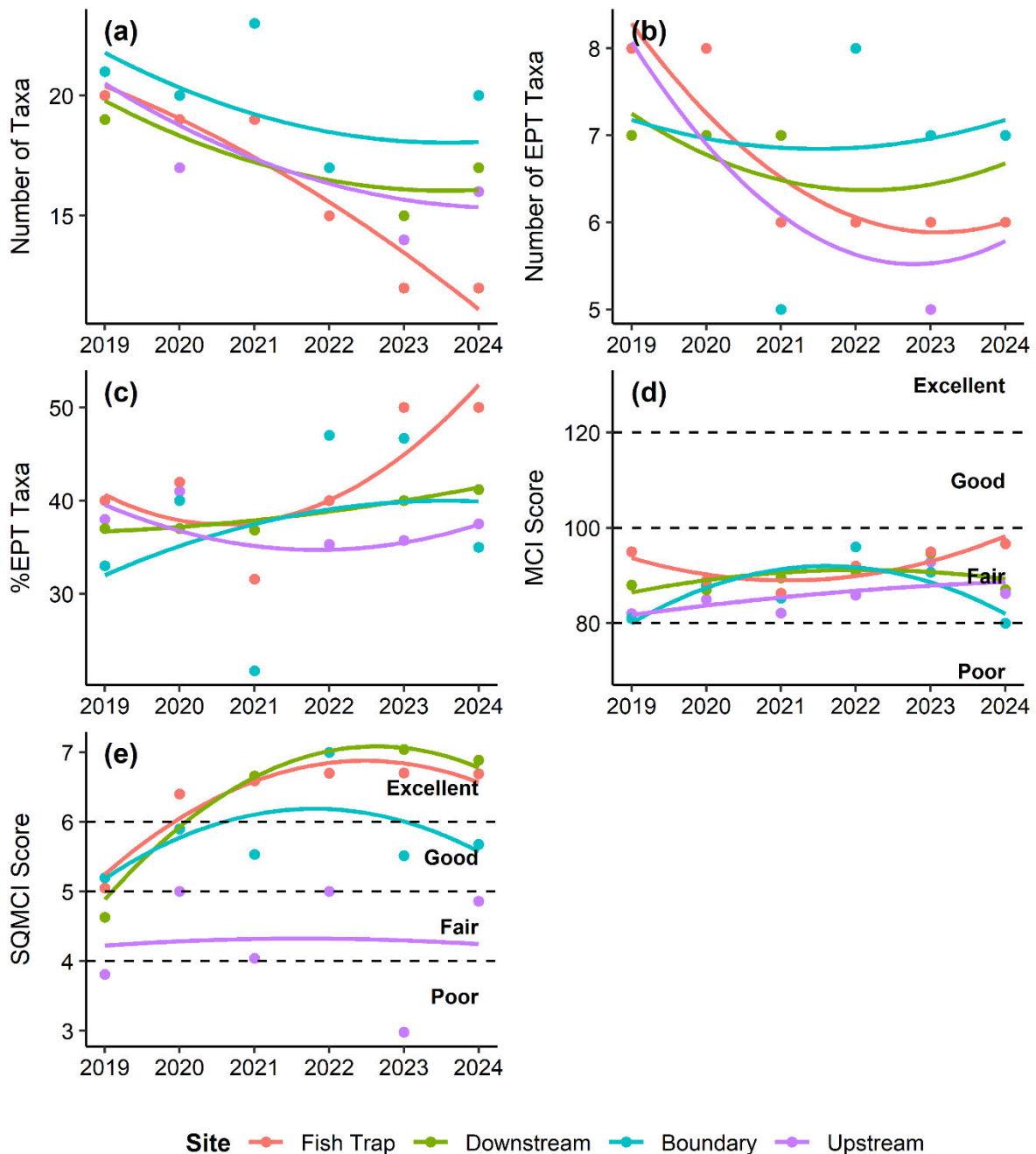
Ultimately, the increased densities of sensitive macroinvertebrate taxa in recent years indicates that development works in Mill Creek have not negatively impacted the local macroinvertebrate communities. On the contrary, it is likely that the recent habitat improvements at Mill Creek have facilitated the development of a healthier freshwater ecosystem overall.

Kōura, freshwater crayfish (*Paranephrops zealandicus*), have historically been present in the Mill Creek catchment. Whilst kōura have not been identified from the Mill Creek macroinvertebrate samples described above, nor in replicate environmental DNA (eDNA) samples collected from the Fish Trap site in May 2023 (from Wilderlab<sup>1</sup>), local iwi have confirmed that they have caught them within the catchment, and Mana Tāhuna Charitable Trust (a kaupapa Māori organisation working in the Lake Hayes catchment) have recently spotted a kōura in the creek (Lauren Christie, Winton, pers. comm.). Grainger *et al.* (2018) lists the conservation status of New Zealand freshwater invertebrates that have been assessed using the New Zealand Threat Classification System (NZTCS); kōura is classified as 'At Risk – Declining'.

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<sup>1</sup> <https://www.wilderlab.co.nz/explore>





**Figure 8: Long-term macroinvertebrate metrics across Mill Creek sites. Regression lines represent a line of best fit but may not accurately reflect trends due to the sample size being too small for reliable statistical analysis and when there is high metrics variability.**



### 3.6 Fish

Fish species recorded from Mill Creek, based on the replicate eDNA samples taken from the Fish Trap site in May 2023, include brown trout (*Salmo trutta*), common bully (*Gobiomorphus cotidianus*), Gollum galaxias (*Galaxias gollumoides*), grass carp (*Ctenopharyngodon idella*) and European perch (*Perca fluviatilis*). eDNA samples from Lake Hayes (in December 2023 and April 2024) found the same species, except for Gollum galaxias, and also found kōaro (*Galaxias brevipinnis*).

The presence of grass carp in the samples is suspected to have come from fish that escaped from authorised releases in contained ponds near Queenstown. The presence of Gollum galaxias is uncertain, given the known distribution of Gollum galaxias does not extend to the Clutha River catchment.

Records from the New Zealand Freshwater Fish Database (NZFFD) indicate the presence of brown trout, common bullies, and kōaro in Mill Creek, with perch, common bully, and brown trout recorded in Lake Hayes.

Common bully is a native species with a conservation status of 'Not Threatened' (Dunn *et al.* 2018). Kōaro is a native species with a conservation status of 'At Risk – Declining' (Dunn *et al.* 2018). Brown trout and perch are both introduced sports fish species.

Only two of these fish species have been found in Mill Creek within the vicinity of Ayrburn Farm: brown trout and kōaro. An electric fishing survey undertaken in Mill Creek by REL in July 2018 found 19 juvenile brown trout (length range 55-115 mm) within the Ayrburn Farm area and two kōaro approximately 300 m further upstream.

Kōaro are one of five migratory galaxiid species that make up the New Zealand whitebait catch and are classified as 'At Risk – Declining' (Dunn *et al.* 2018). Kōaro spawn in late autumn to winter, with eggs laid on bankside partially submerged gravels. Once hatched, the larvae are swept down to lakes or the sea to feed and grow before returning inland as post-larval 'whitebait'. Both juveniles and adults are skilled climbers and will migrate far inland. Kōaro have an affinity for fast-flowing stream sections like riffles.

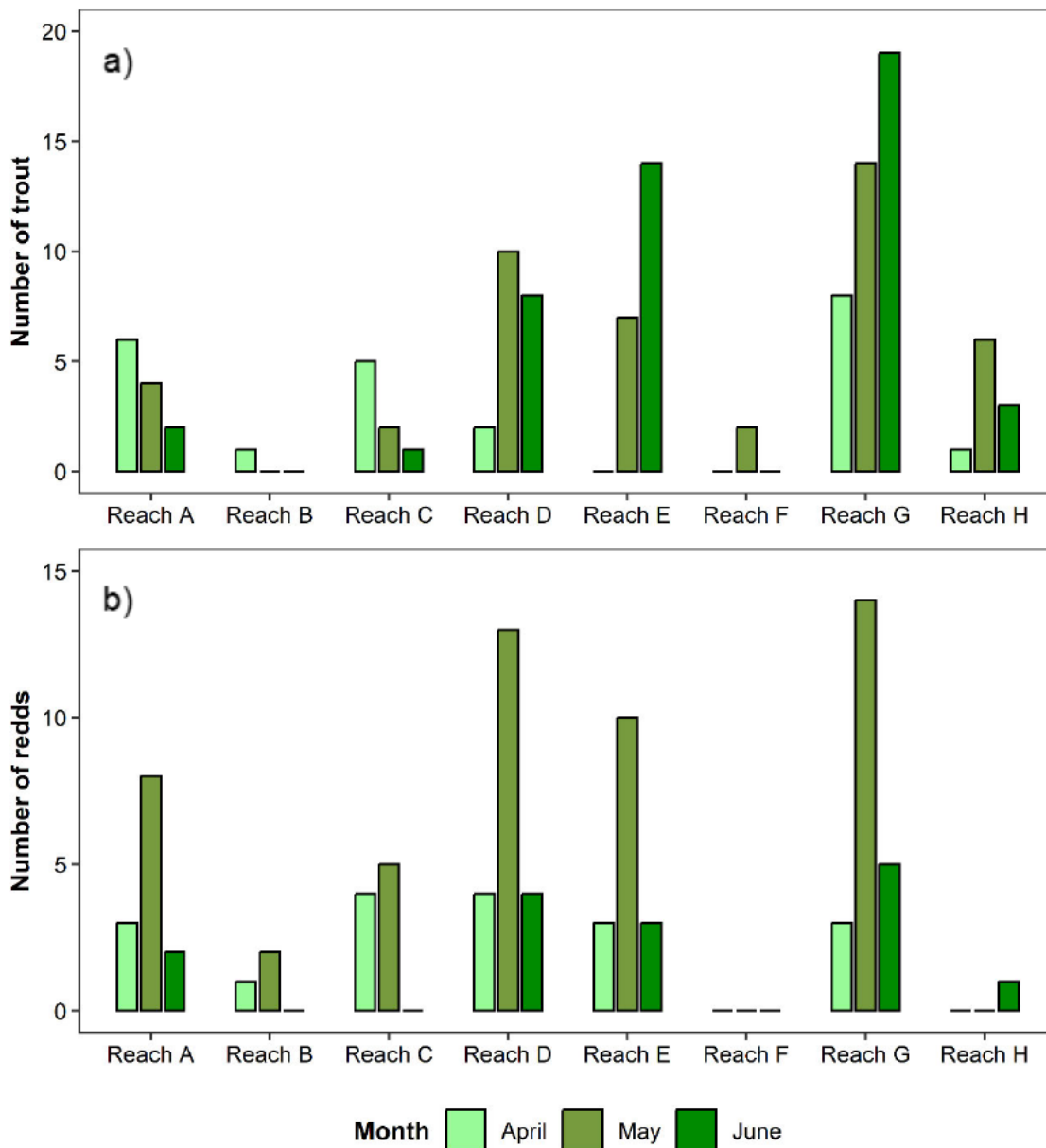
Brown trout are an introduced sports fish species that often outcompetes and predate upon native fish like kōaro. Brown trout spawn in autumn to winter, creating nests (redds) in gravel beds of freshwater streams and rivers, where their eggs incubate and hatch into alevins, then fry, and eventually mature into adults that can live up to 20 years. They inhabit cool, oxygen-rich waters and may migrate between freshwater and coastal environments.

Mill Creek is an important spawning tributary for brown trout. Since 2019, bankside observations of trout have been undertaken in May and June between the Downstream and Upstream sites, with the 2024 sampling also including an event in April. In 2024, a large population of brown trout was observed throughout Mill Creek, particularly in the mid to upper reaches where there was a greater availability of ideal spawning gravels (Figure 9a). The most significant activity was recorded during May and June, coinciding with the upstream migration of trout from Lake Hayes to spawn. Additionally, trout redds were most prominent in the mid reaches (Figure 9b), where ideal spawning gravels were abundant. Redd counts were notably higher in May, suggesting that this was the peak spawning period.

Long-term trout observations have either remained stable but low or have declined in recent years and redd counts have increased in upper-mid catchment reaches. It is important to note that there will be high variability with bankside trout observations used to determine trout population size in Mill Creek due to the high availability of instream refuges (e.g., steep/undercut banks) and the increased cover provided by riparian plantings. While instream refuges and riparian vegetation can obstruct trout observations, these characteristics provide many positive benefits including being ideal habitats for trout and the



broader stream ecosystem. Moreover, although some reaches may appear to not have good spawning conditions, these distinct habitats are likely still valuable refuge, nursery, or feeding grounds for juvenile trout or native fish species that will generally be missed when undertaking bankside surveys.



**Figure 9: Number of brown trout (a) and redds (b) observed across Mill Creek reaches, 2024. Reaches in order from downstream (Reach A) to upstream (Reach H) (see Figure 6).**

### 3.7 Summary

Overall, the current state of the aquatic ecosystem in Mill Creek is of good quality and while several long-term water quality metrics were highest around 2022, most metrics have decreased back to levels recorded in 2019. The exceptions are nitrogen compound





concentrations, which show an increasing trend, however the increasing trend is the same at the Upstream and Downstream sites and is therefore related to wider upper catchment activities and not localised within the development area or related to the construction activities in Waterfall Park.

With more stable and less disturbed instream and riparian habitats, Mill Creek water quality will likely remain more consistently healthy in the long-term. Notably, for the 2024 and 2025 sampling, all water quality metrics met relevant RPW limits except nitrate nitrite nitrogen. Mill Creek supports a healthy macroinvertebrate community and adequate feeding, refuge, and breeding grounds for a relatively large trout population.

## **4.0 Assessment of Actual and Potential Effects**

### **4.1 Project Components**

The specific components of the Screen Hub are described in detail in the Ayrburn Screen Hub Design Report dated 3 June 2025 and in the EMP (Enviroscope 2025).

Construction and operation of the Screen Hub has the potential for adverse effects on ecological values during construction and/or operation. Potential effects are listed below and are discussed in the following sections:

- Sediment discharge during construction.
- Contaminant spills and pest introduction.
- Concrete.
- Stormwater discharge post development.
- Sediment trap construction and maintenance.

### **4.2 Sediment Discharge During Construction**

Without proper erosion and sediment control, sediment discharge is possible during construction of the sediment trap and Ayrburn Screen Hub development, especially through surface runoff during high rainfall and works in and adjacent to Mill Creek. Sediment discharges can affect water quality and downstream macroinvertebrate and fish communities by altering the water's chemical and physical properties and affecting periphyton, a food source for invertebrates. Moreover, increased sedimentation can fill refuges and interstitial gaps between boulders and coarse gravels which are key refuges and spawning grounds for macroinvertebrates and fish.

The EMP (Enviroscope 2025) includes a range of sediment and erosion control methods to manage sediments during construction works, including the use of silt fences along boundaries of the ephemeral watercourse and Mill Creek, the creation of sediment retention ponds, clean water diversion channels, and dirty water diversion channels. Bulk earthworks associated with the sediment trap will be undertaken in spring and summer when periods of fine weather are more frequent.

The use of 'clean water' diversion channels, to capture and divert water from surfaces above the exposed works sites and convey the water around earthworks, should reduce the amount of water moving through disturbed sites. Additionally, the use of 'dirty water' diversion channels, to capture and carry sediment-laden stormwater to sediment retention ponds, should reduce the potential for sediment inputs to waterways. The use of sediment retention ponds, including existing ponds used for previous works associated with the Ayrburn Development, will allow for sediments to settle out from the water column.





Monitoring will also be undertaken at the outlets of the sediment retention ponds (see Section 5.0 of this report) to ensure contributions to Mill Creek are not ecologically harmful.

Ultimately, by following the best-practice erosion and sediment controls and guidelines, as outlined in the EMP, during general construction and the development of the proposed stormwater management system, sediment retention ponds, and sediment trap, adverse effects on the ecological values of the watercourses from the construction activities will be avoided or minimised.

### 4.3 Contaminant Spills and Pest Introduction

The presence of construction machinery on site presents a risk of contaminants (e.g., diesel, lubricants) entering watercourses, with the potential to harm aquatic life. Machinery brought to the site from elsewhere may also spread pest species. While didymo has been historically found in Mill Creek, care should still be applied to prevent further spread throughout Mill Creek and to other local watercourses, especially near waterways following construction.

The EMP (Enviroscope 2025) includes:

- Storage of chemicals and fuels will be as far as practicably possible from waterways and concentrated flows.
- The refuelling of vehicles and plant onsite will occur in a designated refueling bay.
- Wastes shall be stored safely and in an organised manner until recycling, reuse, or disposal offsite.
- Weeds will be treated prior to disturbance of the natural surface, and weed free topsoil will be retained for reuse in site rehabilitation.

Construction activities associated with the Ayrburn Screen Hub, undertaken in accordance with the EMP, will avoid or minimise any adverse effects on the ecological values of the watercourses from the construction activities.

### 4.4 Concrete

Careful management of concrete use during the construction phase is essential to minimise environmental risks, particularly near waterways. Concrete-related activities can pose a significant threat to aquatic ecosystems, as concrete dust, runoff from fresh concrete, and the washing of concrete equipment can introduce harmful contaminants into the water. These contaminants, such as lime, can rapidly increase the pH to highly alkaline levels, compromising water quality and potentially leading to fish kills and other ecological disturbances.

To mitigate these risks, concrete should only be used under controlled conditions. Specifically, all concrete work should be carried out during dry weather and confined to dry riverbeds to prevent materials from entering water systems. Furthermore, strict protocols for handling, washing, and disposing of concrete and related materials should be implemented to further safeguard the surrounding environment. Regular monitoring and contingency plans should also be in place to promptly address any accidental discharges.

The EMP (Enviroscope 2025) includes that all concrete washing is to be undertaken in a designated concrete wash-out pit.

Construction activities associated with the Ayrburn Screen Hub that involve concrete, undertaken in accordance with the EMP, will avoid or minimise any adverse effects on the ecological values of the watercourses from the construction activities.



## 4.5 Stormwater Discharge Post Development

Following the completion of construction activities, future stormwater discharges can contain contaminants (e.g., suspended sediments, oxygen demanding substances, toxicants, and elevated nutrient levels) that can have different water quality attribute values (e.g., temperature, conductivity) to current discharges and therefore have the potential to adversely affect water quality and ecological communities in the receiving water of Mill Creek.

The proposed stormwater management concept for the Ayrburn Development, described in the SMP (CKL 2025), provides for collection of stormwater runoff from roofs, roads, and open space and conveyance in a network system of gardens and wetlands leading to stormwater retention ponds, developed as part of the long-term stormwater management system for the site. Stormwater management for the Screen Hub area includes a network of sediment retention ponds, rain gardens, engineered wetlands, and a treatment pond that will regulate sediment and water run-off from the area. For frequent minor rainfall events, the infiltration of water to ground in the swales, rain gardens, and treatment pond would minimise the discharge of stormwater from the additional impervious surfaces to Mill Creek. It is possible that the lower reaches of the ephemeral tributary will receive less stormwater runoff due to the placement of intercepting diversion channels in the western area of the Screen Hub, which could lead to increased desiccation events.

Although the greater stormwater drainage can contain contaminants, the stormwater retention ponds and other stormwater management features should mitigate these inputs by allowing contaminants to settle. Moreover, in storm events the pond should capture and treat the “first flush” (i.e., the initial runoff from the surface), which typically carries the highest load of contaminants.

## 4.6 Sediment Trap Construction and Maintenance

### Specifications and Construction

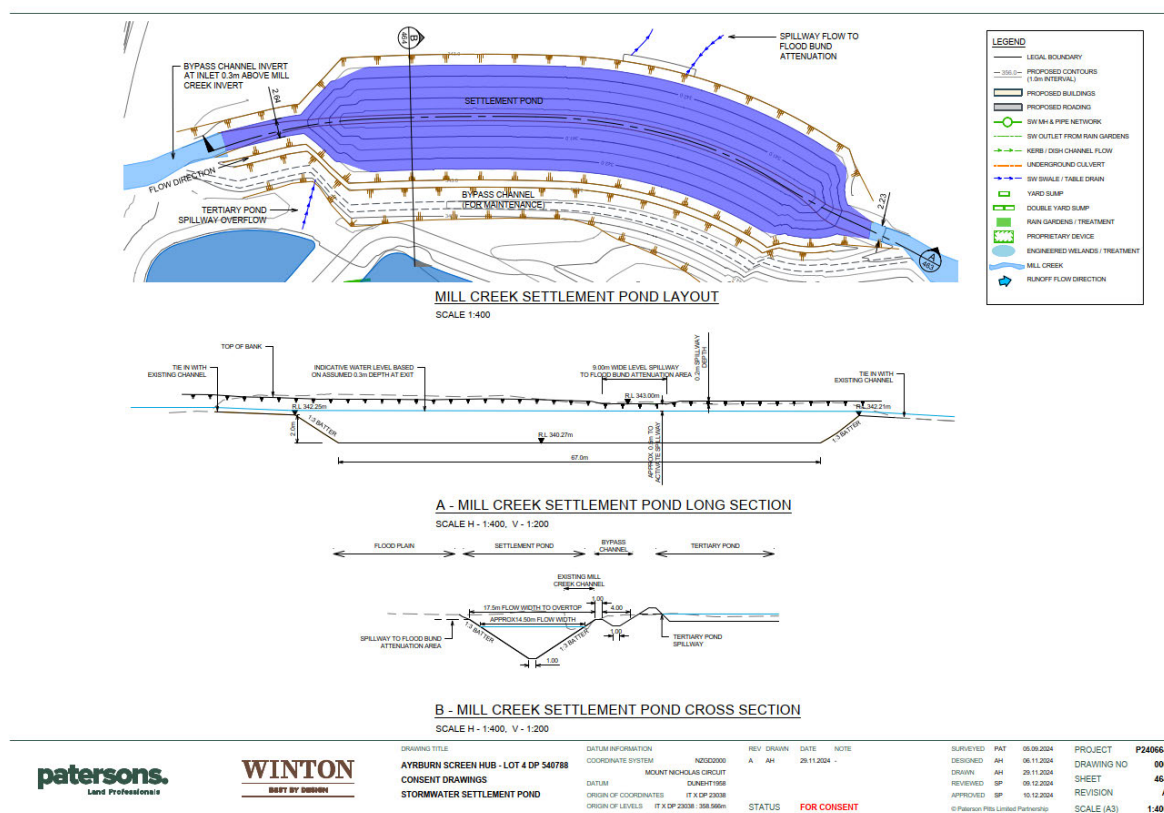
The creation of an in-line sediment trap (50 m long by 12 m wide) in Mill Creek will result in disturbance of the stream bed, the alteration of instream habitat, and will require stream diversions and dewatering. A temporary diversion channel, to convey Mill Creek flows around the sediment trap, will be constructed for temporary use during construction of the trap and during maintenance of the trap.

The construction process for the sediment trap and associated diversion channel is briefly outlined below:

- 1 The temporary diversion channel will be constructed off-line.
- 2 Mill Creek flows will be diverted into the temporary diversion channel.
- 3 Once Mill Creek has been diverted, the existing creek bed will be widened to create the sediment trap. This stage of the construction process will be completed within three months and will be undertaken outside of the brown trout and kōaro spawning season.
- 4 Mill Creek flows will be rediverted back into the original channel alignment, which will now include the sediment trap.
- 5 The temporary diversion channel will remain in place and will be grassed. This channel will then be re-used when cleaning out of the sediment trap is next required.

A schematic of the sediment trap is shown in Figure 10.





**Figure 10: Schematic of the proposed sediment trap. Prepared by Patersons.**

Additional information regarding the construction process, including the use of sediment and erosion control and regarding the use of earth bunds and non-erodible dams, is provided in the EMP (Enviroscope 2025) and discussed below. The construction methodology has been prepared in accordance with Section G4 of the GD05 best-practice methodology for works within a watercourse.

Construction works will be undertaken when Mill Creek is at a low flow level and outside of key spawning periods. Mill Creek will be maintained through its current flow path and kept isolated until construction of the diversion channel is completed. Appendix C shows a drawing of the proposed works.

Prior to dewatering of the section of Mill Creek, and 'livening' of the new temporary diversion channel, fish recovery and relocation must be undertaken with captured fish/invertebrates to be relocated upstream of the diversion. Fish relocation will be completed using a combination of netting and electrofishing, where required, along the length of the diversion.

Once the diversion channel is stabilised and fish relocation is complete, the downstream plug will be opened to allow Mill Creek water to flow up the diversion channel, keeping some water within the channel to reduce problems when the upstream plug is excavated. When the upstream plug is opened, water from Mill Creek will flow into the diversion channel.

Mill Creek will be isolated by the use of non-erodible dams at the upstream and downstream ends of the existing Mill Creek channel. The isolated section of Mill Creek will be drained by pumping, with the water treated via the adjoining sediment retention pond, before it re-enters the live section of Mill Creek downstream. The in-line sediment trap will be constructed as per engineer design, approximately 50 m long x 12 m wide x 2 m deep.



Once the sediment trap has been constructed and is stabilised, the downstream plug will be opened to allow water to flow up the channel whilst keeping some water within the channel to reduce problems when the upstream plug is excavated. The upstream plug will then be opened, allowing water to flow through the sediment trap. Following fish recovery and relocation, the diversion channel will then be drained.

### Effects Assessment and Effects Management

The installation of an in-line sediment trap (see Figure 11 for smaller examples located in the upper Mill Creek catchment) will necessitate the temporary diversion of Mill Creek and the direct disturbance of the streambed. Stream diversions can isolate water containing fish and macroinvertebrates, directly reducing available habitat for aquatic communities, and dewatering of the isolated area can directly remove the animals from the river. Timing of diversions need to be carefully managed to avoid potential spawning periods for fish, as dewatering can isolate spawning areas from flowing water which is required to support development of fish eggs/larvae. Timing of works between 8 January and 15 April will avoid the spawning and migrating periods for fish potentially present in the area, however adults and larval fish could be present. Fish are likely to avoid disturbed areas, but fish salvages will be required during the trap construction when the stream is dewatered and diverted. Following diversion of the stream, and fish salvage, disturbance of the isolated stream bed will be completed under dry conditions to ensure sediment discharges into downstream environments is minimised.

Recovery of brown trout and potentially kōaro from the creek in the vicinity of the proposed sediment trap and from the temporary diversion channel, prior to relevant works commencing, and their relocation to suitable habitats, will be required. Details of the methods to be used for the recovery process are set out in the EMP and are broadly grouped into three methods:

- Netting/trapping prior to dewatering.
- Electric fishing recovery measures on the day the stream is dewatered.
- Where practical, and to minimise injury to fish, preference will be given to encouraging fish to voluntarily leave the watercourse section prior to netting and electric fishing.

Allowing fish to passively vacate a stream during dewatering poses the least risk of injury to fish compared to other methods, but its effectiveness depends on the stream morphology, vegetation density and method of dewatering. Any pools remaining after dewatering will need to be actively fished.

Fish passage will be maintained through the sediment trap, and through the temporary diversion channel when in use, due to the level of the bed of the sediment trap being lower than the bed of the current channel (therefore allowing free flow of water into and out of the sediment trap without any barrier) and the temporary diversion channel will be constructed to convey the required flows without any barriers.

The sediment trap will be positioned between sediment sampling sites 1 and 2 (see Figure 6), areas characterised by a diverse range of sediment sizes and moderate suitability for trout spawning (Figure 7). Additionally, the 'Downstream' macroinvertebrate site, located just below the sediment trap, and the 'Fish Trap' site further downstream, both offer habitats of 'Excellent' quality based on SQMCI scores. The streambed modification for the sediment trap will also take place within a known trout spawning area ('Reach A' in Figure 6), where between 2-6 adult trout and 2-8 redds were observed during the 2024 bankside sports fish surveys (Figure 9). Note that bankside fish surveys do not provide an accurate representation of fish densities or community composition in the direct construction area.



While the project will result in the direct loss of some spawning habitat and high-quality macroinvertebrate habitat, the sediment trap is expected to enhance downstream spawning sites and improve rearing conditions in Lake Hayes by reducing the accumulation of fine sediment. Moreover, sediment traps are known to often be used as pool refuges which provide low flow resting and feeding habitats for freshwater fish. The suitability for the sediment trap to provide fish habitat will also be enhanced by riparian plantings which provide shading and cover; riparian planting will be undertaken along the sides of the sediment trap.

### **Sediment Trap Maintenance**

The maintenance of the sediment trap will involve temporarily diverting Mill Creek through the temporary diversion channel to allow for the excavation of accumulated sediment from within the trap. As with the initial sediment trap installation, fish salvages will be required to safely remove fish first from the sediment trap when water is diverted and later from the temporary diversion channel when water is redirected back over the trap. If a water intake pump is utilised at any stage (e.g., to remove residual water from the trap), the pump inlet will be equipped with an appropriately designed fish screen to prevent the ingress of small fish.

Excavation of trapped sediment is likely to result in a brief sediment flush into Mill Creek when the trap is rewetted, either from residual sediments within the trap or from material that has accumulated in the bypass channel. To mitigate environmental impacts, the trap should be rewetted slowly by rediverting a small portion of the main flow through until the trap is full and running at a similar turbidity to the creek. The excavated sediment must be transported to a suitable disposal site, such as an approved clean fill dump site or an area positioned in a suitable location, sufficiently far from waterways and stabilised with grass to eliminate the risk of runoff and subsequent resuspension into aquatic systems.

Regular monitoring of sediment levels within the trap is essential to ensure timely emptying before it becomes excessively full, which could compromise its ability to effectively capture and settle sediments. It is also important to consider the potential for high-flow events to scour and mobilise previously settled sediment, which could contribute to unintended releases. Therefore, maintenance practices must strike a balance between the frequency of sediment removal (to manage costs and minimize the disturbance caused by sediment plumes) and the effectiveness of the trap in reducing sediment resuspension risks. This balanced approach will ensure the trap remains functional while minimising ecological impacts.

Following construction of the sediment trap, the trap will be cleared within the first 24 months of operation to establish the sediment load captured. Water depths in the sediment trap will be monitored every three months, with the trap to be cleared of sediment when water depths in the middle of the trap average less than 400 mm. Records of sediment volumes removed from the trap will be taken.







**Figure 11: Examples of existing sediment traps located in the upper Mill Creek catchment. Top photo from Friends of Lake Hayes. Bottom photo from Mana Tāhuna Charitable Trust (2023).**



## **Freshwater Fisheries Activities (Fast-track Approvals Act 2024)**

The Fast-track Approvals Act 2024 provides for activities for which an approval would otherwise be required under Freshwater Fisheries Regulations 1983, and additional activities, and differentiates between activities that impede fish passage temporarily or works are completed over a shorter (less than three months) period ('standard freshwater fisheries activities'; SFFA) and activities that impede fish passage more permanently or works are completed over a longer (greater than three months) period ('complex freshwater fisheries activities'; CFFA).

Construction of the sediment trap will not impede fish passage, the works required to construct the sediment trap (i.e., active disturbance of the water body) will not persist for more than three months, and the works will be timed to not occur during the spawning season for fish (i.e., works will be between 8 January and 15 April, which is outside the spawning season for brown trout which is May to August). The construction of the sediment trap therefore constitutes an SFFA.

Of relevance to the sediment trap (structure or fish facility), the following information must be provided: a description of the type of structure, and the dimensions, design and placement of the structure, water flows and operating regime; freshwater species and values present (with particular focus on threatened, data-deficient, and at-risk species as defined in the NZTCS); water quality and quantity in the surrounding habitat; and how the passage of fish will be provided for.

This information has been provided in the sections above.

## **5.0 Monitoring**

Water quality monitoring is included in the EMP (Enviroscope 2025), with the objective to assess whether controlled and uncontrolled discharges meet the required Discharge Criteria, specifically:

- Visual Clarity (mm):  $\geq 100$  mm (As per GD05).
- Turbidity:  $\leq 100$  NTU.
- Total Suspended Sediment:  $\leq 50$  mg/L.
- pH: 5.5 – 8.5.
- Hydrocarbons or tannins: No visible trace.
- Waste: No waste or litter is visible.

Monitoring is to be undertaken at boundaries of sites where any water is flowing, with three sites identified for outlets from the sediment retention ponds, two sites in Mill Creek (upstream and downstream of the sediment trap), and one site in the ephemeral watercourse.

## **6.0 Positive Effects**

The proposed development will result in a change of land use for Ayrburn Farm, with existing land use including a vineyard, carpark, storage area for developments, and open pasture. Associated with this land use change could be a reduction in nutrient loss through leaching to groundwater. The groundwater system below Ayrburn Farm is connected to Lake Hayes, with groundwater from the Mid Mill Creek Aquifer emerging to the surface at Rutherford Road Springs and discharging to the lake from the north-western edge (Rekker 2020).





The overall goal of the Lake Hayes Management Strategy (ORC 1995) is, '*to improve the water quality of Lake Hayes, to achieve a standard suitable for contact recreation year round and to prevent further algal blooms*'. A major concern for the management of the lake has been reducing phosphorus inputs to the lake (ORC 1995, Hydrosphere Research 2017). Phosphorus inputs are a particular concern because phosphorus binds to sediments, typically in the form of phosphate, through adsorption or incorporation into minerals. Under certain conditions, such as low oxygen levels (anoxia) in the bottom waters, phosphorus can be released back into the water column. This release promotes the growth of algae and aquatic plants, leading to eutrophication, a process where excessive nutrients cause algal blooms, reduced water clarity, oxygen depletion, and harm to aquatic ecosystems. Over time, this cycle can degrade water quality and biodiversity in lakes.

The main source of phosphorus to the lake is through surface water, predominantly via Mill Creek. Phosphorus losses from Ayrburn Farm under the land use at the time of an assessment in 2020 were estimated at 4 kg/year (Mudge and Lee 2020). Phosphorus is primarily lost via run-off from the land, particularly during high rainfall events, with the erosion and mobilisation of sediment carrying particulate phosphorus within surface water. Stormwater management at Ayrburn will help reduce sediment inputs to Mill Creek. During high rainfall events the ephemeral watercourse will also transport phosphorus lost from the surrounding land to Mill Creek (and ultimately Lake Hayes). For most of its length, the ephemeral watercourse is currently open and its banks dominated by pasture grasses. Proposed riparian planting in these areas will be beneficial to shade the watercourse and help filter any runoff to the channel.

The proposed land use change to the Screen Hub can be expected to reduce phosphorus and faecal bacteria inputs to surface water through the installation of some hard surfaces (i.e. preventing surface erosion) and the revegetation of the ephemeral watercourse. The stormwater retention pond will also capture and treat the "first flush" runoff, which typically carries the highest sediment load.

Hudson *et al.* (2023) calculated contaminant loads entering Lake Hayes from the Mill Creek catchment and estimated that approximately 81% of the total phosphorus load entering Lake Hayes entered Mill Creek upstream of the Waterfall Park area. Hudson *et al.* (2023) recommended reviewing the mitigation actions proposed by Goeller *et al.* (2020<sup>2</sup>), which looked at diffuse pollution mitigation options for Mill Creek. According to Hudson *et al.* (2023), these mitigation actions included maintaining and restoring existing wetlands and riparian buffers, constructing sediment traps along the main stem of Mill Creek to capture total suspended solids and total phosphorus, livestock exclusion (particularly in the upper catchment), channel restoration in the lower catchment to slow movement of water and reduce bank erosion, re-establishing riparian vegetation and making use of riparian buffer elements such as grass filter strips, mixed vegetation buffers and shrubs and trees. Many of these recommendations, such as riparian vegetation and buffers, channel restoration, and livestock exclusion, have already been implemented as part of the wider Waterfall Park development and are proposed as part of the proposed Screen Hub development.

The proposed sediment trap (50 m long by 12 m wide) in the main stem of Mill Creek is estimated to be able to capture and hold 1,500 tonnes (900 m<sup>3</sup>) of sediment<sup>3</sup>. Calculations on water velocities and sedimentation are discussed in the CKL (2025) SMP. This sediment trap is a further implementation of the mitigation actions proposed by Goeller *et al.* (2020)

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<sup>2</sup> Goeller, B.C., Sukias, J., and Hughes, A. 2020. Scoping of diffuse pollution mitigation options for Mill Creek. NIWA client report prepared for Friends of Lake Hayes Society, Inc., Otago Regional Council, Queenstown Lakes District Council, Department of Conservation, 2020009HN: 71.

<sup>3</sup> From CKL (2025) and personal communication between Friends of Lake Hayes and Winton, January 2025.





and will also provide significant benefits to the lower reaches of Mill Creek and especially to Lake Hayes. Comparatively, a similar existing sediment trap in the upper Mill Creek catchment, which is 30 m in length and 12 m wide, collected approximately 700 tonnes (360 m<sup>3</sup>) of sediment in 11 months<sup>4</sup>. With the majority of phosphorus and sediment inputs entering Mill Creek in the upper catchment, the location of this larger sediment trap at the Ayrburn Farm lower in the Mill Creek catchment will not only mitigate any sediment inputs locally around the site, but also the more substantial inputs originating from the broader catchment. Regular excavation of sediment from the trap is proposed and will be required by consent conditions volunteered by WPD, to ensure the trap remains effective at capturing sediments from the wider catchment and significantly reducing sediment inputs into Lake Hayes.

Overall, the proposed developments associated with the Screen Hub and specifically the addition of a large sediment trap low in the Mill Creek catchment should provide further ecological benefits and improvements to the local area of Mill Creek through to Lake Hayes itself. Notably, these proposed works, in addition to the extensive works stabilising and planting the banks of Mill Creek throughout the Waterfall Park development area, will contribute significantly to reducing sediment and phosphorus loads entering the waterway. The combined impact of these measures is anticipated to mitigate nutrient loss, enhance water quality, and promote ecological health. By addressing key sources of phosphorus and sedimentation, these initiatives align with other initiatives to support the long-term goal of improving the water clarity and ecological balance of Lake Hayes, while also demonstrating a commitment to sustainable land use and environmental stewardship in the region.

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<sup>4</sup> Personal communication between Friends of Lake Hayes and Winton, January 2025.



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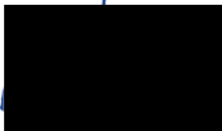


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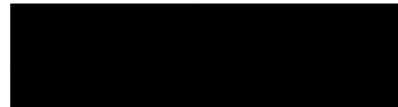
## 8.0 Closure

Sincerely,

**SLR Consulting New Zealand**



**Ben Ludgate**  
Principal Ecologist



**Fleur Tiernan**  
Principal Ecologist





# **Appendix A   Authors’ Qualifications and Experience**

## **Ecology Assessment**

**Ayrburn Screen Hub**

**Waterfall Park Developments Limited**

SLR Project No.: 875.V11822.00003

25 July 2025

The contributing authors, in their capacity as authors of this report, have read and abide by the Environment Court of New Zealand's Code of Conduct for Expert Witnesses Practice Note 2023. Where this report relies on information provided by other experts, this is outlined within the report.

Name & Role	Qualifications	Experience
<b>Ben Ludgate</b> Principal Ecologist Team Manager Christchurch and Dunedin (Ecology & Marine Science)	MSc., Zoology, University of Otago (2001) BSc., Zoology, University of Otago (1998)	Ben has over 22 years of experience as a consultant freshwater ecologist, initially for Ryder Consulting/Environmental that is now part of SLR. Ben specialises in assessments of water quality, periphyton, macrophytes, and macroinvertebrate and fish communities. Ben has managed ecological surveys and assessments across New Zealand for a range of clients including for renewable energy generation, receiving water monitoring in relation to discharges, and land development. Ben also manages the processing of algae, invertebrate, and fish samples in SLR's Dunedin laboratory.
<b>Fleur Tiernan</b> Principal Freshwater Ecologist	MSc. Environmental Engineering MSc. Freshwater Ecology BSc. (Hons) Zoology & Geology	Fleur is a Principal Consultant with over 20 years of experience working with regulatory authorities and consultants in the UK and New Zealand. <ul style="list-style-type: none"> <li>Designing and implementing monitoring programmes for freshwater and coastal environments.</li> <li>Undertaking freshwater assessments, including habitat assessments, stream ecological evaluations, macroinvertebrate monitoring, sediment assessments etc.</li> </ul>





# **Appendix B    Mill Creek Water Quality Site Photographs**

## **Ecology Assessment**

**Ayrburn Screen Hub**

**Waterfall Park Developments Limited**

SLR Project No.: 875.V11822.00003

25 July 2025





**Figure B.1: Mill Creek, 'Fish Trap' monitoring site, 2024.**



**Figure B.2: Mill Creek, 'Downstream' monitoring site, 2024.**







**Figure B.3: Mill Creek, 'Boundary' monitoring site, 2024.**



**Figure B.4: Mill Creek, 'Upstream' monitoring site, 2024.**







# **Appendix C    Mill Creek Macroinvertebrate Community, 2024**

## **Ecology Assessment**

**Ayrburn Screen Hub**

**Waterfall Park Developments Limited**

SLR Project No.: 875.V11822.00003

25 July 2025

**Table C.1: Benthic macroinvertebrate communities in Mill Creek, May 2024. Coded abundance categories from Stark (1998). Sites in order from downstream to upstream.**

ORDER	TAXON	MCI tolerance value	Fish Trap	Downstream	Boundary	Upstream
COLEOPTERA	Elmidae	6	A	A	R	R
CRUSTACEA	Ostracoda	3	R		C	C
DIPTERA	<i>Austrosimulium</i>	3		R	R	R
	Empididae	3			R	R
	<i>Maoridiamesa</i>	3		C	VVA	A
	<i>Mischoderus</i>	4	R			
	Muscidae	3	R		R	R
	Orthoclaadiinae	2		C	R	A
	<i>Paralimnophila</i>	6		R		
	Tanytarsini	3			R	
EPHEMEROPTERA	<i>Deleatidium</i>	8	VVA	VVA	VVA	A
MOLLUSCA	<i>Gyraulus</i>	3		R	C	C
	<i>Physa</i> = <i>Physella</i>	3		R		
	<i>Potamopyrgus</i>	4	A	VA	VA	VA
	Sphaeriidae	3		R	C	
NEMATODA	NEMATODA	3			R	
OLIGOCHAETA	OLIGOCHAETA	1	VA	A	A	A
TRICHOPTERA	<i>Hudsonema</i>	6		A	C	R
	<i>Hydrobiosis</i>	5	C		C	R
	<i>Hydropsyche</i> - <i>Aoteapsyche</i>	4	R	A	VA	R
	<i>Oxyethira</i>	2		R	R	
	<i>Psilochorema</i>	8	C	C	C	R
	<i>Pycnocentria</i>	7	VA	VA	VVA	VA
	<i>Pycnocentrodes</i>	5	A	C		
Number of taxa			12	17	20	16
Number of EPT taxa (incl. Hydroptilidae)			6	7	7	6
Number of EPT taxa (excl. Hydroptilidae)			6	6	6	6
% EPT taxa (incl. Hydroptilidae)			50	41	35	38
% EPT taxa (excl. Hydroptilidae)			50	35	30	38
MCI score			97	87	80	86
SQMCI score			6.7	6.9	5.7	4.9





# **Appendix D   Erosion and Sediment Control Plan Drawing (from EMP)**

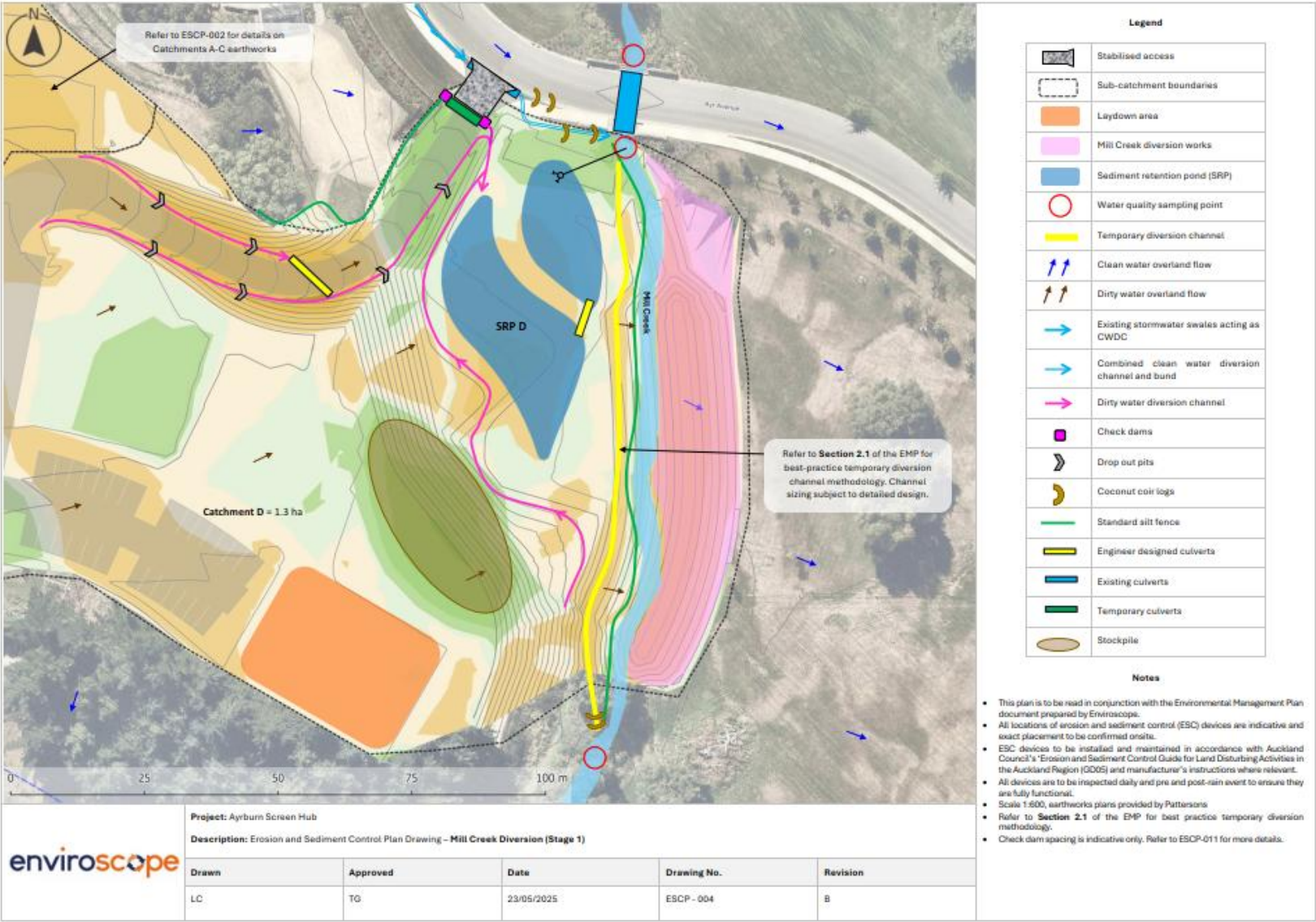
## **Ecology Assessment**

**Ayrburn Screen Hub**

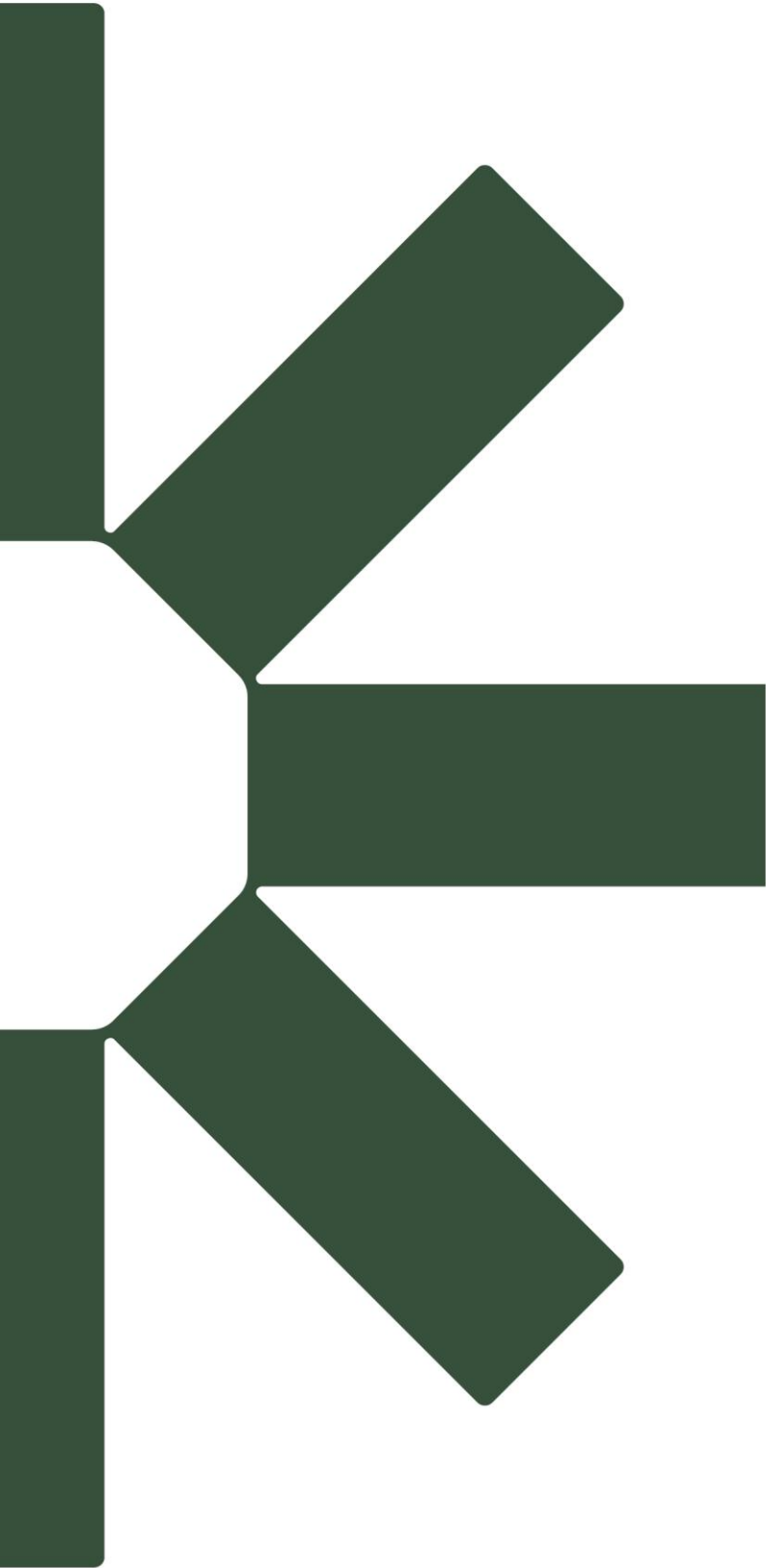
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